



Radiometric Measurements in of Japanese barberry (*Berberis thunbergii* DC.), Boxwood (*Buxus sempervirens* L.) and Gold tassel (*Euonymus japonica* Thunb.) Under Cadmium and Zinc Stress

Nezahat Turfan^a * , Erkan Genç^b 

^a Department of Biology, Science Faculty, Kastamonu University, Kastamonu, Türkiye

^b Central Second School, Kastamonu, Türkiye

*Corresponding Author: nturfan@kastamonu.edu.tr

Received: July 31, 2022 ♦ Accepted: November 28, 2022 ♦ Published Online: December 26, 2022

Abstract: In this study, the effects of Cd and Zn applications on the activity concentration and transfer factors in the Japanese barberry, Boxwood, and Gold tassel leaves were investigated using gamma-ray spectrometry. The mean concentrations (in Bq kg⁻¹) of radionuclides in the studied soil samples were found to be 289.40±32.47 for ²³⁸U, 241.76±27.47 for ²³²Th, 783.63±83.46 for ⁴⁰K, and 31.44±5.63 for ¹³⁷Cs while the respective values in the studied species were 168.6±20.1- 288.8±34.5, 145.9±19.1-250.3±32.4, 434.6±52.2-828.4±99.4, and 16.1±1.8-28.3±3.3. The activity concentrations were found to be at the lowest in the control group and 400 µM Zn for all three species, and at the highest level at 25 µM Cd in general in the species. The order of radionuclides by the highest activity concentrations was ⁴⁰K>²³⁸U>²³²Th>¹³⁷Cs, whereas the order of species was Gold tassel>Boxwood>Japanese barberry. TF (²³²U, ²³²Th, ⁴⁰K, and ¹³⁷Cs) values were found to be between 0.583 and 0.998, between 0.604 and 1.036, between 0.555 and 1.057, and between 0.513 and 0.899. And also, while the order of species by the activity concentration was Gold tassel>Boxwood>Japanese barberry the order of species by the TF values was Boxwood>Gold tassel>Japanese barberry. In conclusion, plants' radionuclide activity concentrations were found to be at the highest level in 25 µM Cd group and at the lowest level in the control group. Considering all the data, it can be stated that a low dose of Cd was effective on the radioactivity concentrations and Gold tassel could be used as the indicator plant in radiation pollution.

Keywords: Cadmium, Zinc, Radioactivity, Transfer factor

Öz: Bu çalışmada, kadın tuzluğu, şimşir ve altuni taflan bitkilerinde Cd ve Zn uygulamalarının radyoaktivite konsantrasyon değişiklikleri ve topraktan yaprağa taşınma faktörü üzerindeki etkileri gama ışını spektrometresi kullanılarak araştırılmıştır. İncelenen toprak örneklerinde radyonuklitlerin ortalama konsantrasyonları (Bq kg⁻¹) ²³⁸U için 289.40±32.47, ²³²Th için 241.76±27.47, ⁴⁰K için 783.63±83.46 ve ¹³⁷Cs için 31.44±5.63 olarak bulunurken, bitki türlerinde bu değerler sırası ile 168.6±20.1- 288.8±34.5, 145.9±19.1-250.3±32.4, 434.6±52.2-828.4±99.4, ve 16.1±1.8-28.3±3.3 (Bq kg⁻¹) olarak bulunmuştur. Türlerde aktivite konsantrasyonları her üç bitki türünde kontrol grubu bitkilerde ve 400 µM Zn dozlarında en düşük ve 25 µM Cd dozunda ise genel olarak en yüksektir. Radyonuklitlerin en yüksek aktivite konsantrasyonlarına göre sıralaması ⁴⁰K>²³⁸U>²³²Th>¹³⁷Cs ve türlerin sıralaması ise Altuni taflan>Şimşir>Kadın tuzluğu olmuştur. Türlerde TF (²³²U, ²³²Th, ⁴⁰K ve ¹³⁷Cs) değerleri sırası ile 0.583-0.998, 0.604-1.036, 0.555-1.057 ve 0.513-0.899 arasında bulunmuştur. TF değerlerine göre radyonuklitlerin sıralaması ⁴⁰K>²³²Th>²³⁸U>¹³⁷Cs ve türlerin sıralaması ise Şimşir>Altuni taflan>Kadın tuzluğu şeklindedir. Sonuç olarak bitkilerde radyonuklit aktivite konsantrasyonları 25 µM Cd dozunda en yüksek, kontrol grubu bitkilerde ise en düşüktür. Tüm veriler göz önünde bulundurulduğunda düşük dozda Cd'in radyoaktivite konsantrasyonlarında etkili olduğu ve Altuni taflanın radyasyon kirliliğinde indikatör bitki olarak kullanılabileceği söylenebilir.

Anahtar Kelimeler: Kadmiyum, Çinko, Radyoaktivite, Transfer faktörü

1. Introduction

As in other organisms, plants are inevitably subjected to radiation effect because the radioactive elements having a very long lifetime have created a natural radiation surface in the ecosystem throughout the history of the world [1]. However, the rapid development of industry, rapid growth of population, industrial and domestic wastes caused by unplanned urbanization, and mining and nuclear energy wastes do cause and have caused radiation pollution in air, water, and soil. Moreover, fossil fuels, nitrogenous fertilizer industry, and synthetic fertilizer technology might cause the release of natural radionuclides into the environment [1, 2]. Natural radiation sources consist of natural radionuclides naturally existing in nature such as ²³⁸U, ²³²Th, and ⁴⁰K and the ²³⁸U and ²³²Th degradation series products (²²⁶Ra and ²²²Rn), while the artificial radionuclides such as ⁹⁰Sr, ¹³⁷Cs, and ¹³¹I are released to the environment through nuclear accidents and nuclear weapon trials [3, 4]. Natural and artificial radionuclides bind to the inorganic matter in soil and sediments through

air and water and accumulate in herbal tissues via the roots. Moreover, radionuclides accumulating in aerosols from the atmosphere might penetrate the plant tissues via leaves and barks [3, 5]. Intake of radionuclides from soil to the plant is defined as transfer factor (TF) and it varies depending on soil characteristics such as pH, clay mineral, Ca, K, and organic matter content [6, 7], leaf characteristics of plants, developmental status of organs, and plant species [8], and climatic parameters such as wind speed, precipitation, and humidity [6-9]. Radionuclides' activity concentrations in plant tissues might vary depending on plant genotype and developmental status of organs, as well as the concentration of radionuclides and their chemical behaviors [10]. In literature, it was emphasized that plants took large amount of ^{40}K and ^{226}Ra , low amount of ^{238}U , and very low amount of ^{232}Th from the soil [7, 11]. In plants, it was reported that the intake of ^{40}K and ^{137}Cs occur Japanese through the same mechanism as fundamental element K that ^{40}K and K^+ were analogous and ^{238}U and ^{226}Ra were analogous to Ca, and that concentrations of ^{40}K and ^{232}U in plant tissues might be higher than those of other radionuclides. Besides that, it is also asserted that application of phosphatic fertilizers increased the ^{238}U activity concentration in soil and plant tissues [12]. Until now, in studies on the effects of heavy metal stress in plants, the changes in the amount of necessary molecules in plant growth and development such as photosynthetic pigments, nitrogenous compounds, carbonaceous compounds, secondary metabolites [13], enzymatic and non-enzymatic defense systems [14, 15] and nutrients have been investigated [16, 17]. And also, radioactivity measurements were performed on organs of many plant species such as leaf, stem, and flower [11, 15, 18], various food sources [19, 20], mushrooms [21], soils [2, 22] and water samples [23] from different regions, in Turkey. However, there is no study carried out on the effects of heavy metal stress on the radionuclide activity concentrations in plant leaves. In the present study, it was aimed to investigate the capacity of Cd and Zn treatments to accumulate ^{238}U , ^{232}Th , ^{40}K , and ^{137}Cs radionuclides in Japanese barberry, Boxwood, and Golden tassel plant species widely grown in parks, gardens, and roadsides in the city center of Kastamonu.

2. Material and Method

In the present study, 2-year-old Japanese barberry (*Berberis thunbergii* DC. var. *atropurpurea* Chenault), Boxwood (*Buxus sempervirens* L. var. *rotundifolia* Baill.), and Golden Tassel (*Euonymus japonicus* Thunb. var. *aureomarginatus* Rehder) plants obtained from Kastamonu Municipality's Department of Parks and Gardens were used. Plants were removed out of the plastic tubes, in which they were grown (S1), and planted into 5L pots containing turf and garden soil (Soil 2; 2:1) and irrigated for 4 weeks by using tap water. Then, the plants were grouped as control, cadmium (Cd: 25 μM and 50 μM - $\text{CdSO}_4\cdot\text{H}_2\text{O}$), and zinc (Zn: 200 μM and 400 μM - ZnCl_2) and they were subjected to metal stress applications by using soil (300 ml) depending on the water retention capacity of soil. The concentrations determined for Cd and Zn were dissolved in Hoagland-Arnon's nutrient solution. While the plants in the control group were given only the nutrient solution, the metal stress application was performed using with the nutrient solution. Metal stress application on plants was performed for 8 weeks (twice a week).

Characteristics of soil samples used in the experiment

pH value of soil samples (S1, S2) was found to range between 6.88 and 6.96 and that of irrigation water was found to be 8.60. Of the soil samples used, K, P, S, Mg, and Ca contents (mg kg^{-1}) were found to vary between 27540- 29681, between 5195-3228, between 3074-2712, between 12950-17580, and between 111700- 27880, respectively (Table 1). Fe, Mn, Cu, Zn, Ni, and Cd contents were found to range between 34960- 38490, between 460.5- 709.2, between 36.8- 37, between 71- 80.7, between 58.8 - 74.80, and between 0.41-3.45 (Table 1).

Table 1. Characteristics of soil mixture used in the experiments

	pH	K	P	S	Mg	Ca	Fe	Mn	Cu	Zn	Ni	Cd
S1	6.88	27540±30	5195±2.8	3074±3	12950±60	111700±100	34960±30	460.5±2.0	36.8±0.7	71±0.7	58.8±20.6	3.45±0.3
S2	6.96	29681±6.5	3228 ±2.4	2712 ±	17580 ±	27880± 32	38490 ±30	709.2 ±1.8	37 ±0.6	80.7±1.4	74.80± 1.1	0.41 ±0.1
W	8.60	2894±4.7	11.96±0.3	-	15951±60	14782.86±60	7.82±0.4	0.322±	20.24±0.4	14.70±0.8	10.23±	1.63±0.1

Preparation of leaf and soil samples for the radioactivity measurements

Leaf samples harvested from the plants were dried in an environment without direct sunlight exposure. The samples were kept in a drying oven at 85°C for 24 hours and then pulverized using a blender. The soil samples used in the experiment were dried at room temperature and then pulverized using the laboratory blender. In order to ensure the homogeneity of samples, they were passed through a sieve with 80 Mesh and left for drying in a drying oven at 85°C temperature for 48 hours.

pH measurements in soil samples

pH values of samples were determined using the method of Gülçür [24]. The samples were kept in 1/2.5 pure water for 24 hours and the pH was measured using a digital pH-meter.

Elemental analysis of soil and leaf samples

Some of the dried soil and leaf samples were used in the elemental analysis in Kastamonu University's Central Research Laboratory by using SPECTRO brand XEPOS model XRF device. Some of samples were put into polyethylene containers with 6 cm diameter and 5 cm height and the lids were closed tightly. In order for samples to reach radioactive balance, they were kept for 1 month [11].

Method for the activity concentrations of radionuclides

Gamma-ray spectrometry was performed with FoodGuard-1 3 x 3-inch NaI (Tl) model radiation detector (ORTEC, Oak Ridge, USA) in the Central Research Laboratory of Kastamonu University. The ground leaves were placed into plastic boxes having a diameter of 8 cm and a height of 8 cm and designed to fit the geometry of the detector. Then, the boxes were tightly closed and kept for 1 month. Thus, the formation of radioactive equilibrium between ^{238}U and ^{232}Th and their decay products was allowed and the samples were prepared for counting. The detector was calibrated before the analysis. To analyze the spectra collected in computer memory, the channel corresponding to the input energy must be known. Thus, the types of radioactive nuclei present in the sample can be found. To accomplish the energy calibration, a standard source(s) consisting of nuclei with previous energies is needed. Standard point sources including the peaks of ^{109}Cd , ^{57}Co , ^{133}Ba , ^{22}Na , ^{137}Cs , ^{54}Mn , and ^{60}Co , with energies ranging between 80 and 1400 keV were used for the calibration. After the calibration, each sample was counted in the gamma spectrometer for 50000 sec. Activities of radionuclides obtained at end of the measurements were determined using the following equation:

$$\text{Activity} = \frac{\text{Net area}}{\text{Counting time} \times \text{Sample amount} \times \text{Abundance} \times \text{Yield}} \quad (1)$$

The net areas under the peaks were calculated by subtracting the background from the total area. The radioactivity concentrations of ^{238}U , ^{232}Th , ^{40}K , and ^{137}Cs in the samples were determined by making use of the gamma peaks of natural radionuclides, which were the degradation products of these radionuclides. After determining the activity concentrations of ^{238}U , ^{232}Th , and ^{40}K , the activity concentration of ^{137}Cs isotope in the samples was also determined. The activity concentrations of radionuclides (^{238}U , ^{232}Th , ^{40}K , and ^{137}Cs) were expressed as Bq kg⁻¹ dry weight.

Calculation of Transfer Factor

The rate of radionuclides, which are present in the soil, to transfer from the plant tissues is named transfer factor (TF). Using the equation given below, TF values were calculated with the mean radionuclide activity concentrations found in the leaves of Red barberry, Boxwood, and Gold tassel plants and the soil samples used in growing the plants [11, 25].

$$\text{TF} = \text{Activity concentration of plant (Bq kg}^{-1}\text{)} / \text{Activity concentration of soil (Bq kg}^{-1}\text{)} \quad (2)$$

3. Result and Discussion

Plants illustrate an important link in the transport and distribution of radionuclides and other pollutants in the environment and are often considered Japanese biomonitors of atmospheric pollution [3, 6, 7, 15]. Naturally occurring and fallout radionuclides were investigated in samples of Japanese barberry, Boxwood, and Golden tassel plants.

Changes in ^{238}U , ^{232}Th , ^{40}K , and ^{137}Cs activity in soil samples

The elemental contents of soil samples used in growing the plants are presented in Table 1. The mean values found in the soil samples were 289.40±32.47 for ^{238}U activity concentration, 241.76±27.47 for ^{232}Th activity concentration, 783.63±83.46 for ^{40}K activity concentration, and 31.44±5.63 for ^{137}Cs activity concentration (Table 2). The ^{238}U , ^{232}Th , ^{40}K , and ^{137}Cs activity concentrations of the tap water used in the experiment were 1.48±0.1, 1.17±0.1, 2.65±0.3, and 0.42±0.04, respectively (Table 2). Study results were found to overlap with the activity concentrations reported by Kaya et al. [22] for the soil samples collected from different regions of Gümüşhane province. Researchers found the ^{232}Th , ^{40}K , and ^{137}Cs activity concentrations in the soil samples to range between 9.7±1.15 and 32.52±2.65, between 236.83±7.53 and 889.65±17.63, and between 7.63±1.26 and 39.44±8.57. In another study, Adesiji & Ademola [25] found the ^{238}U , ^{232}Th , and ^{40}K activity concentrations of soil samples they used in growing corn plant to be in the range of 242.13 ± 429.10-2763.90 ± 2345.77, 15294.77 ± 6924.46-26211.90 ± 7178.22, and 374.01 ± 590.51-5008.18 ± 2427.165 Bqkg⁻¹, respectively, and those values are much higher than the results achieved in the present study. ^{40}K activity

concentration was similar to the value reported by Bilgici Cengiz et al. [2-20] (245.6±34.6 Bqkg⁻¹ and 814.2±35.7 Bqkg⁻¹) but ²³²Th (22.2±6.8-44.6±7.5 Bqkg⁻¹) activity concentration was much lower than the value found in the present study.

Table 2. Radioactivity concentration changes in the soil samples used in the experiment

	²³⁸ U	²³² Th	⁴⁰ K	¹³⁷ Cs
Soil 1	236.35±26.20	196.87±16.60	678.64±68.52	22.44±2.70
Soil 2	322.45±38.73	286.64±38.33	876.62±98.40	38.44±8.56
Mean	289.40±32.47	241.76±27.47	783.63±83.46	31.44±5.63
Water	1.48±0.1	1.17±0.1	2.65±0.3	0.42±0.04

¹³⁷Cs activity concentration found in the present study was confirmed by the results achieved by Lamarque et al. [26] (0-5 cm: 61-280 Bqkg⁻¹; 10-15 cm: 14-224 Bq kg⁻¹). But Absar et al. [27] reported the ²³²Th, ⁴⁰K, and ¹³⁷Cs activity concentrations in soil to be in the ranges of 50 ±19-65± 21, 245 ±30-635 ±35, and ¹³⁷Cs 3±1-9±1, respectively. In the present study, ²³²Th and ¹³⁷Cs values were lower than those values but ⁴⁰K was found to be in a similar range.

Radionuclide activity concentration changes in plant samples

The activity concentrations for ²³⁸U, ²³²Th, ⁴⁰K, and ¹³⁷Cs radionuclides found in leaves of Japanese barberry, Boxwood, and Gold tassel plants subjected to Cd and Zn application are presented in Table 3. Given the results, although the activity concentrations of those radionuclides varied by the species and concentration, radioactivity concentrations were found to be higher than in control for all three plants (Table 3).

²³⁸U activity concentration changes and TF values in plants subjected to Cd and Zn

²³⁸U, a natural radionuclide, exists in nature generally in form of uranium minerals with elements such as Ca, Mg, and P. Since it has low solubility in soil solution, its intake by the plants is also at a low level. However, since its chemical behavior is similar to that of Ca, it was reported to have positive effects on metabolic reactions, in which Ca is effective [6, 12, 26]. In the present study, ²³⁸U activity concentrations found in leaves of Red barberry, Boxwood, and Gold tassel were found to be 168.6±20.1-223.7±26.4 1 Bq kg⁻¹, 171.0±20.6-265.9±31.7 1 Bq kg⁻¹, and 176.5±21.2-288.8±34.5 Bq kg⁻¹ (Table 3). In comparison to the control group, the highest level of ²³⁸U activity concentration was found at 25 µM Cd dose in all three species. The lowest activity concentration was found in the control group plants. The second-highest activity concentration was achieved at 50 µM Cd dose for Red barberry and Gold tassel leaves and 200 µM Cd dose for Boxwood leaves (Table 3). Among the plant species, the highest ²³⁸U activity concentrations were found in Gold tassel leaves and with Cd doses (288.8±34.5; 270.1±32.1 Bq kg⁻¹), whereas the lowest activity concentration was found in the control group samples of Red barberry leaves (168.6±20.1 Bq kg⁻¹). ²³⁸U accumulation capacities of plants were found to be Gold tassel>Boxwood>Red barberry. In these plants, TF (²³⁸U) value was reported to be 0.583-0.773 for Red barberry, 0.591-0.919 for Boxwood, and 0.610-0.998 for Gold tassel. In comparison to the control group, the highest TF value was achieved at 25 µM Cd dose for all three plants. TF values reached the maximum levels in Red barberry and Gold tassel leaves with Cd doses and in Boxwood leaves with 25 µM Cd and 200 µM Zn doses (Table 3).

Table 3. Effects of Cd (25 µM and 50 µM) and Zn (200 µM and 400 µM) applications on ²³⁸U and ²³²Th activity concentrations and TF changes in Red barberry, Boxwood, and Gold leaves (Bq kg⁻¹)

Plant	Group	²³⁸ U	²³² Th	²³⁸ U	²³² Th
Japanese barberry	Control	168.6±20.1	148.0±19.1	0.583	0.612
	25 µM Cd	223.7±26.4	232.5±30.1	0.773	0.962
	50 µM Cd	214.5±25.3	173.4±22.4	0.742	0.718
	200 µM Zn	198.2±23.4	218.1±28.1	0.685	0.903
	400 µM Zn	184.7±22.1	168.5±21.7	0.638	0.697
Boxwood	Control	171.0±20.6	151.9±19.6	0.591	0.628
	25 µM Cd	265.9±31.7	244.2±31.4	0.919	1.010
	50 µM Cd	217.8±26.2	179.4±23.3	0.753	0.742
	200 µM Zn	245.4±29.4	211.9±27.4	0.848	0.877
	400 µM Zn	200.9±24.3	167.7±21.6	0.695	0.694
Gold tassel	Control	176.5±21.2	145.9±19.1	0.610	0.604
	25 µM Cd	288.8±34.5	250.3±32.4	0.998	1.036
	50 µM Cd	270.1±32.1	211.7±27.3	0.933	0.876
	200 µM Zn	259.1±31.5	222.0±28.6	0.896	0.919
	400 µM Zn	201.4±24.7	161.6±21.0	0.696	0.669

The ^{238}U activity concentration found in the present study was higher in comparison to the results reported in the literature. Examining several tree species and epiphyte plants, Manigandan et al. [4] reported the mean ^{238}U activity concentration to range between 9.6 ± 0.4 and 11.4 ± 0.4 and TF value to range between 0.249 and 0.313. However, Tshivhase et al. [28] reported the ^{238}U activity concentration to be 31.36 ± 9.40 , 0.02 ± 0.01 , and 0.16 ± 0.14 Bq kg^{-1} and TF value to be 0.19 ± 0.06 , 0.307 ± 0.89 , and 0.11 ± 0.01 Bq kg^{-1} , respectively. In the present study, the finding that ^{238}U activity concentration in experimental groups was higher than in the control group was related to the P concentration in soil and low Cd and Zn doses stimulating the ^{238}U absorption of species. In literature, it was reported that the mean P concentration in soil was 500-800 mg kg^{-1} , that P values in soils in Turkey ranged between 146.2 and 3125 mg kg^{-1} , and that the P concentration required for plant was 0.3-3 kg ha^{-1} [29, 30].

^{232}Th activity concentration changes in plants treated with Cd and Zn

^{232}Th activity concentration was found to be within the ranges of 148.0 ± 19.1 - 232.5 ± 30.1 Bq kg^{-1} in Japanese barberry leaves, 151.9 ± 19.6 - 244.2 ± 31.4 Bq kg^{-1} in Boxwood leaves, and 145.9 ± 19.1 - 250.3 ± 32.4 Bq kg^{-1} in Gold tassel leaves. In the control group plants, the highest activity concentration was achieved in Boxwood leaves and in Gold tassel. ^{232}Th activity concentrations that were the highest in comparison to the control group were observed at 25 μM Cd and 200 μM Zn doses for all three species. Among the plant species, the highest ^{232}Th activity concentration was found in Gold tassel leaves (250.3 ± 32.4 Bq kg^{-1}) and the lowest one in Gold tassel control group (145.9 ± 19.1 Bq kg^{-1}). In the samples, TF(^{232}Th) values were in the ranges of 0.612-0.962 for Japanese barberry leaves, 0.628-1.010 for Boxwood leaves, and 0.604-1.036 for Gold tassel leaves. TF values were generally at the highest levels in low Cd and Zn doses (Table 3). ^{232}Th activity concentration changes achieved in the present study overlapped with the results reported by Adesiji & Ademola [25]. Examining two different corn leaves which they grew in two different soil samples, researchers reported the ^{232}Th activity concentration to vary between 238.05 ± 64.64 and 826.37 ± 1182.03 Bq kg^{-1} , activity concentration in soil samples to range between 1776.08 ± 4164.89 and 26211.90 ± 7178.22 Bq kg^{-1} , and TF value to range between 0.02 ± 1.27 and 0.08 ± 3.70 . However, Chakraborty et al. [31] examining grass and Bilgici Cengiz & Çağlar [32] analyzing 45 wheat flour samples reported ^{232}Th activity concentration values that were much lower than in the present study. Similarly, Absar et al. [27] reported the ^{232}Th activity concentration of tea plant leaves to be 2.4 ± 0.5 - 5.8 ± 0.9 Bq kg^{-1} and that of soil to be 50 ± 13 - 63 ± 5 Bq kg^{-1} , whereas TF (^{232}Th) was found to be 0.05 ± 0.04 . Bilgici Cengic & Çağlar [20] analyzed various herbs widely used in the Eastern Anatolian region and reported the ^{232}Th activity concentration to be 55.99 ± 4.32 Bq kg^{-1} and TF value to be 0.88.

^{40}K activity concentration changes in plants treated with Cd and Zn

^{40}K activity concentration was found to be 434.6 ± 52.2 - 536.2 ± 64.3 Bq kg^{-1} in Japanese barberry leaves, 529.8 ± 63.6 - 828.4 ± 99.4 Bq kg^{-1} in Boxwood leaves, and 534.9 ± 64.2 - 821.4 ± 98.6 Bq kg^{-1} in Gold tassel leaves. In comparison to the control group, the highest activity concentration was found to be 828.4 ± 99.4 in Boxwood leaves treated with 25 μM Cd, followed by Gold tassel leaves treated with 200 μM Zn (821.4 ± 98.6 Bq kg^{-1}). The lowest activity concentration was found in the control group Japanese barberry leaves, followed by Japanese barberry leaves treated with 400 μM Zn (Table 4). Similar to ^{232}Th activity concentration, ^{40}K activity concentration reached the highest levels with low Cd and Zn doses. Among the plants, the highest activity concentration was found in ^{40}K and the order of plants was found to be Boxwood > Gold tassel > Japanese barberry (Table 3). TF(^{40}K) values were found to range between 0.555 and 0.685 in Japanese barberry leaves, between 0.676 and 1.057 in Boxwood leaves, and between 0.683 and 1.048 in Gold tassel leaves. TF(^{40}K) was found to be high in low Cd and Zn treatments in comparison to the control group and other treatments and the highest value was found in Boxwood leaves. Besides that, the order of species by the TF (^{40}K) values was Gold tassel > Boxwood > Japanese barberry (Table 3). The ^{40}K activity concentrations found in plants are in corroboration with the literature. ^{40}K is a naturally rich radionuclide in plants. The fact that we found high concentrations of ^{40}K in the leaves of sample plants is not surprising given that plants obtain their nutrients and water through root uptake from the soil, in which there are high ^{40}K concentrations. The amount of potassium in plants is high because of its essential role in most physiological processes needed to maintain plant growth and development. Potassium has an important role in photosynthesis, translocation of starches and sugars, plant-water relations, protein synthesis, activation of plant enzymes, resistance to plant diseases [8, 29]. Similar results were reported in the studies examining the herbaceous and woody plants [7, 15, 27]. Manigandan et al. [3] reported the ^{40}K activity concentration in some plant species grown in rain forests of India to vary between 160.4 ± 12.3 and 206.4 ± 13.4 Bq kg^{-1} and TF value to vary between 0.802 and 0.954. Similar to the present study, Shayeb et al. [33] determined the ^{40}K activity concentration in date samples to be 181 ± 17 Bq kg^{-1} , the activity concentration in soil samples to be 329 ± 87 Bq kg^{-1} , and TF value to be 0.51 ± 2.0 . In another study examining the medicinal plants used in traditional medicine in Thailand, the mean ^{40}K activity concentration was found to be 610 ± 260 Bq kg^{-1} and TF value to be 2.0 ± 1.4 and it was determined that the activity concentration was at a higher level in leaves in comparison to flowers and stem Saenboonruang et al. [7]. In their study carried out on forests in Southwestern Serbia region, Hadrović et al. [34] reported the ^{40}K activity concentration of evergreen species to be 102 ± 25 Bq kg^{-1} , that of non-evergreen species to be 140 ± 26 Bq kg^{-1} , and that of samples from the soil, where the plants were grown, to be 62 ± 5 - 970 ± 60 Bq kg^{-1} . TF (^{40}K) was found to be 0.022-0.22 in non-evergreen species and 0.007-0.19 in coniferous species. Much higher level of ^{40}K activity concentrations in soil and plant tissues in comparison to other

radionuclides was related to the chemical behaviors of ^{40}K and essential element K^+ . Researchers reported that the intakes of K^+ and ^{40}K occur Japanese through similar mechanisms and their roles in metabolic reactions were also the same [8, 29, 35].

^{137}Cs activity concentration changes in plants treated with Cd and Zn

The ^{137}Cs activity concentration changes found in the plant samples were in the ranges 16.1 ± 1.8 - 26.2 ± 3.1 Bq kg^{-1} in Japanese barberry leaves, 17.7 ± 2.3 - 26.3 ± 3.2 Bq kg^{-1} in Boxwood leaves, and 17.3 ± 2.1 - 28.3 ± 3.3 Bq kg^{-1} in Gold tassel leaves (Table 4). The highest activity concentration was found at 200 μM Zn dose in Golden Tassel (28.3 ± 3.3 Bq kg^{-1}) leaves, followed by 25 μM Cd dose in Japanese barberry (26.3 ± 3.2 Bq kg^{-1}) and Gold tassel leaves (26.3 ± 3.2 Bq kg^{-1}). ^{137}Cs activity concentration was found to be higher in Boxwood and Gold tassel leaves at 25 μM Cd and 200 μM Zn doses, whereas it was high in Japanese barberry leaves at Cd (25-50 μM) doses. The lowest activity concentration in plants was found in ^{137}Cs radionuclide and the order of plants by this parameter was found to be Gold tassel > Boxwood > Japanese barberry (Table 4).

TF (^{137}Cs) values were found to be between 0.550 and 0.899 in Japanese barberry, 0.563 and 0.836 in Boxwood, and 0.513 and 0.835 in Gold tassel leaves. TF (^{137}Cs) was found to be high at low Cd and Zn doses in the first two species and at Cd doses in Gold tassel leaves. The order of species by TF value was Japanese barberry > Boxwood > Gold tassel (Table 4). ^{137}Cs activity concentration data were in corroboration with the literature. In previous studies, the lowest activity concentration in soil and plant organs was reported to belong to ^{137}Cs [36]. Lamarque et al. [26] monitored the seasonal activity concentration changes of ^{137}Cs in *Fagus sylvatica* and *Picea abies* grown in forests, which were polluted because of the Chernobyl disaster, in Franche-Comté region in Northeastern France. Researchers determined that ^{137}Cs activity concentration in soil samples ranged between 61 and 280 Bq kg^{-1} (0-5 cm) and between 14 and 224 Bq kg^{-1} (10-15 cm), that TF values in leaves varied seasonally, and that TF (^{137}Cs) was 0.0074 for *F. sylvatica* and 0.0179 for *P. Abies*. Researchers also reported that there was no direct relationship between cesium activity in soil and cesium activity in plant organs and that it might be because the intake of ^{137}Cs might have occurred Japanese through roots from the soil and through leaves from aerosols in the air. ^{137}Cs activity concentrations in soil and grass in Bangladesh were reported to be 0.17 ± 0.02 Bq kg^{-1} and 2.41 ± 0.18 Bq kg^{-1} , respectively, whereas TF value was found to be 0.061 [31]. Shayeb et al. [34] compared Japanese the ^{137}Cs activity concentrations in date and soil samples collected from different regions of Saudi Arabia and they reported the activity in soil to be 10.2 ± 2.1 Bq kg^{-1} and the activity in date samples to be below the limit of detector. In a study carried out using tea leaves, ^{137}Cs activity concentration was found to be <0.4 Bq kg^{-1} in the leaves and 3 ± 1 - 10 ± 1 Bq kg^{-1} in the soils, whereas TF (^{137}Cs) was found to be below the limit of detector [4]. Hadrović et al. [34] examining the ^{137}Cs activity concentration in forests of Southwestern Serbia reported the ^{137}Cs activity concentration to be 4.9 ± 7.1 Bq kg^{-1} in some non-evergreen species and 5.9 ± 4.8 Bq kg^{-1} in evergreen samples.

Table 4. Effects of Cd (25 μM and 50 μM) and Zn (200 μM and 400 μM) treatments on ^{40}K and ^{137}Cs activity concentration and TF changes in Red barberry, Boxwood, and Gold leaves (Bq kg^{-1})

Plant	Group	^{40}K	^{137}Cs	^{40}K	^{137}Cs
Red barberry	Control	434.6 ± 52.2	16.1 ± 1.8	0.555	0.550
	25 μM Cd	536.2 ± 64.3	26.2 ± 3.1	0.685	0.836
	50 μM Cd	492.9 ± 59.1	19.1 ± 2.3	0.629	0.835
	200 μM Zn	520.3 ± 62.4	18.7 ± 2.4	0.664	0.899
	400 μM Zn	454.6 ± 54.6	18.1 ± 2.2	0.580	0.630
Boxwood	Control	529.8 ± 63.6	17.7 ± 2.3	0.676	0.563
	25 μM Cd	828.4 ± 99.4	26.3 ± 3.2	1.057	0.836
	50 μM Cd	643.9 ± 77.3	20.2 ± 2.4	0.822	0.641
	200 μM Zn	745.1 ± 89.4	23.7 ± 2.6	0.951	0.753
	400 μM Zn	581.7 ± 69.8	19.6 ± 2.5	0.743	0.624
Gold tassel	Control	534.9 ± 64.2	17.3 ± 2.1	0.683	0.513
	25 μM Cd	804.8 ± 96.6	26.3 ± 3.2	1.027	0.835
	50 μM Cd	792.6 ± 95.1	26.2 ± 3.2	1.012	0.608
	200 μM Zn	821.4 ± 98.6	28.3 ± 3.3	1.048	0.595
	400 μM Zn	584.3 ± 70.1	19.8 ± 2.5	0.746	0.575

TF (^{137}Cs) was found to be 5.2 in non-evergreen trees and in the range between 0.021 and 0.18 in coniferous species and activity concentration and TF values were found to be at the highest in leaves. Researchers reported that ^{137}Cs intake of plants occur Japanese through the same mechanism as K^+ intake and it accumulated more in leaves as with the K^+ . Moreover, it was claimed that the plants with larger leaf surface area genes need K^+ element more because of higher transpiration, stomal conductivity, and photosynthetic activity and, thus, more ^{40}K and ^{137}Cs might accumulate in the

plants having larger leaf surface area [8, 36]. Depending on the activity concentrations of radionuclides and the changes in TF, the order of plants by the (1) activity concentration was Gold tassel>Boxwood>Japanese barberry and the order by TF values was Boxwood>Gold tassel>Japanese barberry and that (2) 25 μM Cd and 200 μM Zn doses yielded the highest radionuclide activity concentration and TF values. It suggests that, regarding the order of species by TF and activity concentrations, leaf characteristics were also important as well as the factor genotype. The larger leaves of Gold tassel in comparison to other two species might increase the competition for radionuclide absorption from both soil and air. Boxwood leaves have also larger surface areas when compared to Japanese barberry leaves. High 40K and 238U activity concentrations in both species confirm this conclusion. The plants having more aboveground volume have higher transpiration, hydrolytic resistance, and stoma activity and these plants necessitate more K^+ and Ca^+ elements. K^+ plays important roles in stoma movements, as well as controlling the events of osmosis and turgor [29], while Ca has a specific importance in strengthening the cell wall [37, 38]. In literature, it was reported that 40K was analogous to essential element K^+ and 238U was analogous to Ca^+ [26, 36, 38]. The higher activity concentrations and TF values at lower doses were related to the possibility that low doses of (Cd-Zn) metals might stimulate the radionuclide absorption. It was stated that Cd was a very mobile element and, thus, rate of its transfer from soil to plant and its speed of transfer within the plant were high [37, 38, 39].

4. Conclusions

In the present study, in which the effects of Cd and Zn treatments on 238U, 232Th, 40K, and 137Cs activity concentrations and TF values of Japanese barberry, Boxwood, and Gold tassel plants were examined, it was revealed that activity concentrations varied depending on plant species, metal species, and concentration. In all three species, the radionuclide activity concentrations were found to be at the lowest levels in the control group and 400 μM Zn groups, whereas the 25 μM Cd dose generally yielded the highest level. 238U and 232Th activity concentrations in Gold tassel (25 μM Cd), 40K activity concentration Boxwood (25 μM Cd), and 137Cs activity concentration in Gold tassel (400 μM Zn) were found to be the highest ones in comparison to the control and other groups. Among the control group plants, the lowest activity concentrations were found in Japanese barberry (23U, 40K, 137Cs) and Gold tassel (232Th) leaves, whereas the order of radionuclides by the highest activity concentrations was 40K>238Uz>232Th>137C and that of species by the highest radionuclide activity concentration was Gold tassel>Boxwood>Japanese barberry. Similar to the activity concentration results, TF values of species were found to be at lower levels in control group plants. The lowest TF values were found in Japanese barberry leaves for TF (238U) and TF (40K) and in Gold tassel leaves for TF (232Th) and TF (137C). The highest TF (238), TF (232Th), and TF (40K) were obtained at 25 μM Cd dose and the highest TF (137Cs) was achieved at 200 μM Zn dose. The order of radionuclides by the highest TF values 40K>232Th>238U>137Cs and that of species was Boxwood>Gold tassel>Japanese barberry. Given the results obtained, it can be stated that low doses of Cd and Zn might increase the radioactivity concentrations and that Gold tassel and Boxwood plants could be used as an indicator regarding the radiation pollution.

Competing Interest / Conflict of Interest

The authors declare that they have no competing interests.

Author Contribution

We declare that all Authors equally contribute.

Availability of data and material:

The datasets obtained from this study are available from the corresponding author on reasonable request.

Competing interests: The authors declare no competing interests.

Funding: There is no financial support and commercial support.

5. References

- [1] Isinkaralar, K. (2022). Some atmospheric trace metals deposition in selected trees as a possible biomonitor. *Romanian Biotechnological*, 27(1), 3227-3236.
- [2] Bilgici Cengiz, G., & Çağlar, İ. (2019). Determination of natural radioactivity concentrations of some fertilizers used in Eastern Anatolia of Turkey. *Caucasian Journal of Science*, 6(2), 147-155.
- [3] Manigandan, P.K., & Chandar Shekar, B. (2014). Uptake of some radionuclides by woody plants growing in the rainforest of Western Ghats in India. *Journal of Environment Radioactivity*, 130,63-67.

- [4] Manigandan, P.K., & Chandar Shekar, B. (2015). Leaves of Woody Plants As Bio-Indicators Of Radionuclides In Forest Ecosystems. *Journal of Radioanalytical and Nuclear Chemistry*, 303, 911-917.
- [5] Kılıç, Ö., Belivermiş, M., Topcuoğlu, S., Çotuk, Y., Coşkun, M., Çayır, A., & Küçer, R. (2008). Radioactivity concentrations and dose assessment in surface soil samples from East and South of Marmara region Turkey. *Radiation Protection Dosimetry*, 128 (3), 324-330.
- [6] Kumar, A., Singhal, R.K., Preetha, J., Rupali, K., Narayanan, U., Suresh, S., Mishra, M.K., & Ranade, A.K. (2008). Impact of tropical ecosystem on the migrational behaviour of K-40, Cs-137, Th-232 U-238 in perennial plants. *Water, Air, and Soil Pollution* 192(1-4), 293-302.
- [7] Saenboonruang, K., Phonchanthuek, E., & Prasandee, K. (2018). Soil-to-Plant Transfer Factors of Natural Radionuclides (226Ra and 40K) in Selected Thai Medicinal Plants. *Journal of Environmental Radioactivity*, 184-185, 1-5.
- [8] Bréda, N. (2008). Leaf Area Index. *Encycl. Ecol.* 2148-2154.
- [9] Doi, T., Masumoto, K., Toyoda, A., Tanaka, A., Shibata, Y., Hirose K. (2013). Anthropogenic radionuclides in the atmosphere observed at Tsukuba: Characteristics of the radionuclides derived from Fukushima. *Journal of Environmental Radioactivity*, 122,55-62.
- [10] Chandrashekara, K., & Somashekarappa, H. (2015). Soil to plant transfer factors of radionuclides in *Ficus racemosa* (L.): A medicinal plant. *International Research Journal of Biological Sciences*, 4(9), 43-47.
- [11] Özden, S., & Aközcan, S. (2020). Determination of radionuclide transfer in sunflower on agricultural lands in Kırklareli. *Kırklareli University Journal of Engineering and Science*, 6(2),153-16.
- [12] Bramki, A., Ramdhane, M., Benrachi, F. (2018). Natural radioelement concentrations in fertilizers and the soil of the Mila Region of Algeria. *Journal of Radiation Research and Applied Sciences*, (11),49-55.
- [13] Chen, S., Wang, Q., Lu, H, Li, J., Yang, D., Liu, J., Yan, C. (2019). Phenolic Metabolism and Related Heavy Metal Tolerance Mechanism in *Kandelia obovata* under Cd and Zn Stress. *Ecotoxicology And Environmental Safety*, 169,134-143.
- [14] Kandziora-Ciupa, M., Ciepał, R., Nadgo'rska-Socha, A., & Barczyk, G. (2016). Accumulation of heavy metals and antioxidant responses in *Pinus sylvestris* L. Needles in Polluted and non-Polluted Sites. *Ecotoxicology*, 25,70-981.
- [15] Turfan, N., Kunaz, A., & Sarıyıldız, T. (2021). Effect of air pollution on element profile and radioactive compounds in six tree species. *Tree and Forest* 2(2),82-92.
- [16] Saleh, E.A.A., & Işınkaralar, Ö. (2022). Analysis of trace elements accumulation in some landscape plants as an indicator of pollution in an urban environment: Case of Ankara. *Kastamonu University Journal of Engineering and Sciences*, 8(1)1-5.
- [17] Karakeçi, H., Kaya, Ö.F., Çelik, T.H. (2021). An Investigation on heavy metal pb, zn, cu, ni and cd accumulation in leaves of *Robinia Pseudoacacia* L. "Umbraculifera" arising from motor vehicles. *Kastamonu University Journal of Engineering and Sciences* 7(2)114-126.
- [18] Kılıç, Ö. (2012). Biomonitoring of 137Cs, 40K, 232Th, and 238U using oak bark in Belgrade Forest, Istanbul, Turkey. *Nuclear Technolog and Radiation Protection*, 27(2),137-143.
- [19] Kurnaz A., &Turfan N (2017). The effect of different storage conditions on the radiometric and element content of the Taşköprü garlic (*Allium sativum*). *Turkish Journal of Agriculture-Food Science and Technology*, 5(4), 373-379.
- [20] Bilgici Cengiz, G., & Çağlar, İ. (2022). Transfer Factors of Natural Radionuclides from Soil to Medicinal Plants Used by Local People in Eastern Anatolia, Turkey. *International Journal of Environment and Geoinformatics*, 9(2), 039-044.
- [21] Pekşen, A., Kurnaz, A., Turfan, N., & Kibar, B. (2021). Determination of radioactivity levels in different mushroom species from Turkey. *Yuzuncu Yil University Journal of Agricultural Science*, 31(1), 30-41.
- [22] Kaya, S., Karabıdak, S.M., & Çevik, U. (2015). Determination of natural (226ra, 232ta and 40k) and artificial (137cs) radioactivity concentrations in soil and moss samples collected from around Gümüşhane. *Gümüşhane University Journal of Science and Technology*, 5 (1), 24-33.
- [23] Kurnaz, A., Turhan, Ş., & Alzaridi, F.M.N.S. (2021). Radiological and physicochemical properties of drinking waters consumed in the Western Black Sea Region of Turkey. *Journal of Radioanalytical Nuclear Chemistry*, 328, 805-814.
- [24] Gülçur, F. (1974). Physical and chemical analysis methods of soil. *Istanbul University Faculty of Forestry Publications No: 201, Istanbul*.
- [25] Adesiji, N.E., & Ademola, J.A. (2019). Soil-to-maize Transfer Factor of Natural Radionuclides in a Tropical Ecosystem of Nigeria. *Nigeria Journal of Pure & Applied Physics*, 9(1),6-10.
- [26] Lamarque, S., Lucot, E., & Badot, P.M. (2005). Soil-plant transfer of radiocaesium in weakly contaminated forest ecosystems. *Radioprotection*, 1(40),407-412.

- [27] Absar, N., Abedin, J., Rahman, M.M., Miah, M.M.H., Siddique, N., Kamal, M., Chowdhury, M.I., Sulieman, A., Faruque, M.R.I., Khandaker, M.U., Bradley, D.A., & Alsubaie, A. (2021). Radionuclides transfer from soil to tea leaves and estimation of committed effective dose to the Bangladesh Populace. *Life*, 11, 282.
- [28] Tshivhase, V.M., Njinga, R.L., Mathuthu, M., & Dlamini, T.C. (2015). Transfer rates of ^{238}U and ^{232}Th for *E. globulus*, *A. mearnsii*, *H. filipendula* and hazardous effects of the usage of medicinal plants from around gold mine dump environs. *International Journal of Environmental Research and Public Health* 12(12), 15782-15793.
- [29] Marschner, H., & Marschner, P. (2012). Marschner's mineral nutrition of higher plants, 3rd ed. (San Diego, Elsevier Academic Press), 1-651.
- [30] Mordoğan, N., Ceylan, Ş., Delibacak, S., Çakıcı, H., Günen, E., Pekcan, T., & Çolak, B. (2012). Effect of organic fertilization to nutrients content in soils cultivated olives. *Journal of Adnan Menderes University Agricultural Faculty*, 10(3),7-13.
- [31] Chakraborty, S.R., Azim, R., Rahman, A.R., & Sarker, R. (2013). Radioactivity concentrations in soil and transfer factors of radionuclides from soil to grass and plants in the Chittagong City of Bangladesh. *Journal of Physical Science*, 24 (1), 95.
- [32] Bilgici Cengiz, G., & Çağlar, İ. (2021). Determination of the natural radioactivity distribution and consumption effective dose rate ff cereal crops in Ardahan Province, Turkey. *Journal of Scientific Reports-A*, 47, 174-183.
- [33] Shayeb MA, Alharbi T, Baloch MA, Rahman Alsamhan OA (2017). Transfer Factors for Natural Radioactivity into Date Palm Pits. *Journal of Environmenta Radioactivity*, 167,75-79.
- [34] Hadrović, S.H., Čeliković I.T., Krneta Nikolić, J.D., Rajačić M.M., & Todorović, D.J. (2021). Radionuclides' content in forest ecosystem located in southwestern part of Serbia. *Nuclear Technology and Radiation Protection*, 36(2),192-196.
- [35] Kant, S., Kant, P., & Kafkafi, U. (2005). Potassium uptake by higher plants: from field application to membrane transport. *Acta Agronomica Hungarica*, 53, 443-459.
- [36] Wilkins, K.A., Matthus, E., Swarbreck, S. M., & Davies, J.M. (2016). Calcium Mediated Abiotic Stress Signaling in Roots. *Frontie in Olant Sciences*, 7,1296.
- [37] Thury, Y., & Van Hess, M. (2008). Evolution of pH, organic matter and ^{226}Ra /calcium partitioning in U-mining debris following revegetation with pine trees. *Science of the Total Environment*, 393,111-117.
- [38] Salt, D.E., & Wagner, R.J. (1993). Cadmium transport across the tonoplast of vesicles from oat roots evidence for a $\text{Cd}^{2+}/\text{H}^{+}$ antiport activity. *Journal of Biological Chemistry*, 268, 12297-12307.
- [39] Kabata-Pendias, &A., Pendias, H. (2001). Trace elements in soils and plants. 3rd Edition, CRC Press, Boca Raton, 403 p.