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Seasonal analysis of Reyhanlı Dam Lake zooplankton fauna

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This study was conducted seasonally in 2022 and 2023 in Reyhanlı Dam Lake (Hatay, Türkiye), which was established in 2020. Zooplankton fauna and main water quality characteristics were investigated. The average values of the water quality parameters (temperature, dissolved oxygen, conductivity and pH) measured in the dam lake were determined to be of first-class water quality. A total of 67 species were identified, of which fifty-four (54) species were Rotifera (80.60%), 7 species were Copepoda (10.45%) and 6 species were Cladocera (8.95%). Brachionidae was the most abundant with 14 species, Chydoridae represented by two species and Cyclopidae were represented by 3 species. According to the frequency index, 3 species ($F \geq 76\%$) were classified as constant and 11 species ($75\% > F \geq 51\%$) were classified as common. Of the 67 species recorded, only 29 species were very abundant (●) and abundant (○) in the different seasons. The highest species richness was recorded in the summer, with 54 species.

INTRODUCTION

Due to the lack of taxonomic studies, it is not possible to take the necessary measures for sustainable management and especially the protection of ecosystems. In Türkiye, which has a very rich fauna composition, unfortunately, there is still not enough taxonomic information on most living groups (Bozkurt and Genç, 2018). Türkiye, which has a very rich inland water potential, has many rivers, lakes and dams with an area of about one million hectares, which are negatively affected by increasing environmental degradation and colonisation.

Many reservoirs have been built for years for drinking water supply, irrigation, flood control and energy production (Yüksel, 2015), but as a result of population growth and industrialization, reservoirs are at risk of eutrophication, which can lead to a loss of biodiversity and disruption of the balance of the food chain (Brito et al., 2011). Therefore, limnological and biological factors in reservoirs should be

studied and assessed, and the results should be used to improve water quality. Biotic and abiotic factors in the reservoirs can influence the diversity, density, biomass and spatio-temporal distribution of zooplankton species (Dorak et al., 2019).

Some aquatic organisms feed on zooplankton only at a certain stage of their life, particularly the larval stage, while many other species feed on zooplankton throughout their lives (Sales, 2011). This explains the close relationship between the diversity and abundance of zooplankton and the productivity of the aquatic environment (Brun et al., 2019). They play an important role in the aquatic environment as most zooplankton organisms such as copepods, cladocerans and rotifers feed on phytoplankton and rapidly convert plants into animal protein (Svanberg et al., 2022).

Although zooplankton is an important component of the food chain, some species are considered good indicators of eutrophication, pollution and water quality due to their

sensitivity to environmental changes (Ismail and Adnan, 2016). Zooplankton abundance and diversity, which are closely linked to water quality characteristics, increase and decrease with the trophic status of inland waters (İpek Alış and Saler, 2016), making studies of inland zooplankton increasingly important. Characterization of the zooplankton fauna of Türkiye, will contribute to a comprehensive understanding of Türkiye's biodiversity. In order to use these inland water resources efficiently, it is important to know the inland waters, aquatic organisms and their distribution in Türkiye.

This study was the first to investigate zooplankton biodiversity and main water quality characteristics in Reyhanlı Dam Lake (Hatay), a new reservoir. It also serves as an example for future studies.

MATERIALS AND METHODS

The study was conducted seasonally between May 2022 and December 2023 in the Reyhanlı Reservoir (36°20'27 "N 36°33'53 "E) in Hatay province (Figure 1). The Reyhanlı Reservoir (Hatay, Türkiye) has a total height of 29 m and a volume of 480 million m³. It was built in 2020 on the Afrin and Karasu rivers for the purpose of agricultural irrigation. Although 5846.40 ha of agricultural land was created for irrigation purposes, it also creates new habitats for many aquatic animals and birds (Şimşek, 2022).

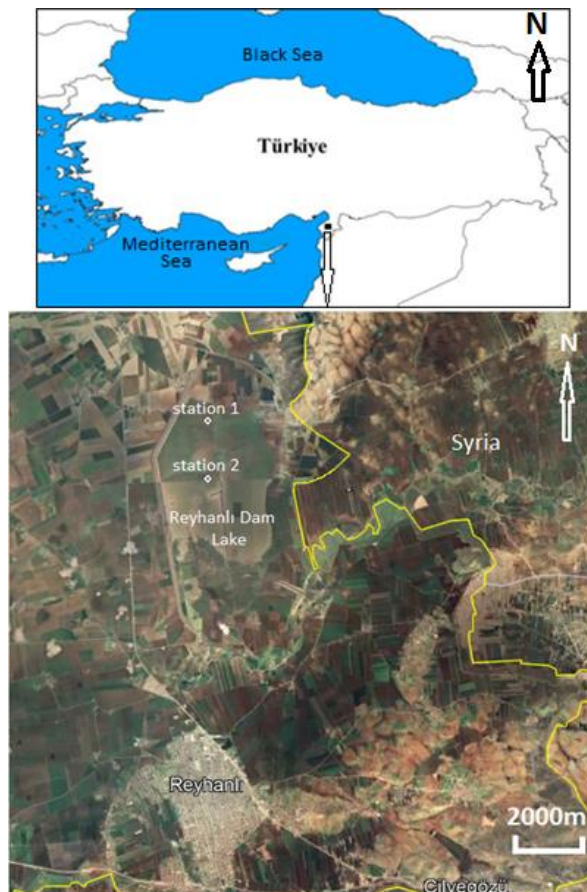


Figure 1. Reyhanlı Dam Lake and sampling stations

Zooplankton samples were collected seasonally at two different stations using a plankton net with a diameter of 0.30 m and a mesh size of 60 µm with horizontal and vertical hauls. Vertical hauls were performed ten times from the bottom to the surface, while horizontal hauls were performed from the water surface for 20 minutes at about 2 km/h with a motorboat. Macrophytes (*Elodea* sp. and *Ceratophyllum* sp.) collected from the northern and southern shores of the reservoir were brought to the laboratory in bags, washed in containers and filtered through the plankton net. All zooplankton samples were fixed in 4% formalin in the laboratory. Dissolved oxygen, water temperature, pH and conductivity were determined in situ using digital meters (oxygen and temperature: YSI Model 52 oxygen meter; pH: YSI 600 pH meter; conductivity: YSI Model 30 salinometer). The first sampling station is closer to the dam lake shore and at a depth of approximately 4-5 m, while the second station is closer to the dam midpoint and at a depth of 12-15 m, and its distance to the crest is around 500 m. Zooplankton species were examined and identified using an inverted microscope and a binocular microscope (Olympus CH40). Approximately 20 cc of each sample was sub-sampled to identify species and examined in a petri. This process was done at least 3 times in order to identify all species present. Soyer's (1970) frequency index (F %) was used to express the frequency of zooplankton species identified from plankton and plant samples in the study area. The results were determined as constant (F ≥ 76%), common (76% > F ≥ 51%), occasionally (51% > F ≥ 26%) and rare (F < 25%). Individuals belonging to the species were not counted, but their abundance was determined visually from their density in the petri dish. This was expressed as rare, few, abundant and most abundant, regardless of the Soyer index (Table 2). The specimens were identified according to Rylov (1963), Borutsky (1964), Scourfield and Harding (1966), Dussart (1967), Dussart (1969), Damian Georgescu (1970), Smirnov (1974), Negrea (1983), Apostolov and Marinov (1988), Reddy (1994), Segers (1995), Karaytug (1999) and Holynska et al. (2003).

RESULTS

Water temperature varied between 11.05±0.21°C (winter) and 25.00±0.28°C (summer), with a mean value of 19.89±5.80°C (Table 1). Dissolved oxygen ranged from 8.05±0.07 mg L⁻¹ (summer) to 9.55±0.21 mg L⁻¹ (winter), with a mean value of 8.75±0.59 mg L⁻¹ (Table 1). The conductivity value ranged from 139±1.41 µS cm⁻¹ (winter) to 235±7.07 µS cm⁻¹ (fall), with a mean value of 187±43.23 µS cm⁻¹ (Table 1). The minimum, maximum and mean pH values were 7.40±0.14 (summer), 8.65±0.07 (spring) and 8.18±0.51 (Table 1), respectively.

Table 1. Main water quality parameters (mean±SD)

Parameters	Spring	Summer	Fall	Winter	Annual Mean
Temperature (°C)	23.55±0.49	25.00±0.28	19.95±0.21	11.05±0.21	19.89±5.80
DO (mg L ⁻¹)	8.60±0.14	8.05±0.07	8.80±0.14	9.55±0.21	8.75±0.59
EC (µS cm ⁻¹)	156±5.66	218±3.54	235±7.07	139±1.41	187±43.23
pH	8.65±0.07	8.25±0.07	7.40±0.14	8.40±0.14	8.18±0.51

A total of 67 species were identified, of which fifty-four (54) species were Rotifera (80.60 %), 7 species were Copepoda (10.45 %) and 6 species were Cladocera (8.95 %) (Table 2). Among the rotifers, in which a total of 17 families were identified, the family Brachionidae was the most abundant with 14 species, followed by Lecanidae with 13 species. On the other hand, Trichocercidae, Mytilinidae, Asplanchnidae, Gastropodidae, Conochilidae, Dicranophoridae, Hexarthridae and Epiphanidae were represented with one species each. Of the Cladocera, five families were recorded, with the Chydoridae represented by two species, while the other families (Bosminidae, Moinidae, Macrothricidae and Ilyocryptidae) were represented by one species each (Table 2). Among the 5 families of Copepoda, the Cyclopidae were represented by 3 species and the other families (Ameiridae, Laophontidae, Ergasilidae and Lernaecidae) by one species each (Table 2).

The most common rotifers recorded in all seasons were, *Brachionus angularis*, *B. bidentatus*, *B. calyciflorus*, *B. caudatus*, *B. falcatus*, *B. quadridentatus*, *Keratella cochlearis*, *K. tecta*, *K. tropica*, *Platyias quadricornis*, *Lecane bulla*, *L. closterocerca*, *L. hamata*, *L. scutata*, *Polyarthra dolichoptera*, *P. vulgaris*, *Synchaeta stylata*, *Trichocerca pusilla*, *Asplanchna sieboldi*, *Collotheca pelagica*, *Pompholyx sulcata* and *Dichranophorus epicharis* followed by *Anuraeopsis coelata*, *B. budapestinensis*, *B. urceolaris*, *L. aculeate*, *L. flexilis*, *L. inermis*, *L. stenroosi*, *C. adriatica*, *C. colurus*, *L. patella*, *C. forficula*, *C. gibba*, *R. neptunia*, *D. hertzogi*, *C. mutabilis*, *A. ovalis* and *F. longiseta* (found in 3 seasons) (Table 2).

Among the Cladocera, *M. micrura*, which was recorded in 4 seasons, had the largest distribution habitat, followed by *B. longirostris*, *M. laticornis* and *I. sordidus* (recorded in 3 seasons). On the other hand, *A. robustus* and *O. mohammed* had the largest distribution area among copepods (found in 4 seasons) (Table 2). Some zooplankton species had a limited distribution and were only detected in one season: *L. furcata*, *L. hornemanni*, *L. luna*, *L. pyriformis*, *Lepadella rhomboids*, *T. ruttneri*, *Testudinella patina*, *C. unicornis* (Rotifera) and *M. rubellus* (Copepoda) (Table 2). The amount of zooplankton was generally not very abundant, with only some species

occasionally reaching high densities. Of the 66 species recorded, only 29 species were very abundant (●) and abundant (○) in the different seasons, while the other species were fewer in number.

Most zooplankton species were recorded in summer with 54 species. This was followed by spring with 53 species, fall with 52 species and winter with 37 species. In terms of abundance, 7 species were very abundant (●) in spring, 7 species in summer and 3 species in fall. The seasonally very abundant (●) species were *B. angularis* (spring, summer and fall), *K. tecta*, *P. vulgaris* (spring, summer), *K. cochlearis*, *P. dolichoptera*, *P. sulcata*, *B. longirostris* (spring), *B. caudatus*, *K. tropica*, *S. stylata*, *N. hibernica* (summer), *C. pelagica* and *M. micrura* (fall) (Table 2). However, zooplankton was abundant (○) with 12 species in each season in spring, summer and fall and 5 species in winter (Table 2). The most abundant (○) species were *R. neptunia* (spring, summer, fall); *B. bidentatus*, *S. stylata*, *A. robustus* (spring, fall); *B. budapestinensis*, *P. dolichoptera* (summer, fall); *B. falcatus* (summer, winter); *B. quadridentatus*, *C. forficula*, *A. sieboldi*, *C. pelagica*, *C. vicinus* (spring, summer); *K. cochlearis*, *C. gibba*, *D. epicharis* (summer); *K. quadrata*, *N. hibernica*, *F. longiseta* (spring); *K. tropica*, *B. caudatus*, *B. longirostris*, *M. micrura* (winter); *P. vulgaris*, *D. hertzogi*, *M. ventralis*, *P. sulcata*, *A. costata*, *M. laticornis* (fall).

Species not found in aquatic plant samples but only in plankton samples; *A. coelata*, *B. angularis*, *B. budapestinensis*, *B. calyciflorus*, *B. falcatus*, *K. cochlearis*, *K. quadrata*, *K. tecta*, *P. vulgaris*, *S. stylata*, *A. sieboldi*, *P. sulcata*, *F. opoliensis*, *P. tentaculatus*, *M. micrura*, *C. vicinus*, *E. sieboldi*, *L. cyprinacea*. In contrast, some zooplankton species (*L. aculeate*, *L. flexilis*, *L. hamata*, *L. inermis*, *L. stenroosi*, *C. colurus*, *L. patella*, *C. forficula*, *E. najas*, *D. hertzogi*, *C. pelagica*, *C. mutabilis*, *D. epicharis*, *M. laticornis* and *I. sordidus*) were only found in aquatic plant samples. Some species (*L. furcata*, *L. hornemanni*, *L. pyriformis*, *T. ruttneri*, *T. patina*, *C. unicornis*, *M. rubellus*) were not assessed because they were found only once during the study, and others were found in both plants and plankton samples.

Table 2. List and abundance of zooplankton species in the dam lake (pl: plankton, op: on the plant. -: Absent, *: very few (rare 25-0.0%), +: few (occasionally 50-26%), o: abundant (Common 75-51%), ●: very abundant (Constant 100-76%)

Species	Sampling Time								F%
	18.05.2022		10.06.2023		03.09.2023		17.12.2023		
	pl	op	pl	op	pl	op	pl	op	
Rotifera									
Brachionidae									
<i>Anuraeopsis coelata</i> de Beauchamp, 1932	*	–	–	–	*	–	*	–	38
<i>Brachionus angularis</i> Gosse, 1851	●	–	●	–	●	–	*	–	50
<i>Brachionus bidentatus</i> Anderson, 1889		o	–	+	*	o	*	+	75
<i>Brachionus budapestinensis</i> Daday, 1885	–	–	o	–	o	–	*	–	38
<i>Brachionus calyciflorus</i> Pallas, 1766	*	–	+	–	*	–	+	–	50
<i>Brachionus caudatus</i> Barrois & Daday, 1894	–	*	●	*	+	–	o	–	63
<i>Brachionus falcatulus</i> Zacharias, 1898	*	–	o	–	+	–	o	–	50
<i>Brachionus quadridentatus</i> Hermann, 1783	+	o	o	o	*	+	+	*	100
<i>Brachionus urceolaris</i> Müller, 1773	–	–	+	–	+	–	–	+	38
<i>Keratella cochlearis</i> (Gosse, 1851)	●	–	o	–	+	–	+	–	50
<i>Keratella quadrata</i> (Müller, 1786)	o	–	+	–	–	–	–	–	25
<i>Keratella tecta</i> (Gosse, 1851)	●	–	●	–	+	–	*	–	50
<i>Keratella tropica</i> (Apstein, 1907)	+	+	●	*	*	*	o	+	100
<i>Platyias quadricornis</i> (Ehrenberg, 1832)	–	+	+	–	–	*	*	–	50
Lecanidae									
<i>Lecane aculeata</i> (Jabuski, 1912)	–	*	–	*	–	–	–	+	38
<i>Lecane bulla</i> (Gosse, 1886)	–	*	–	*	*	+	–	*	63
<i>Lecane closterocerca</i> (Schmarda, 1859)	–	+	–	+	*	–	*	*	63
<i>Lecane flexilis</i> (Gosse, 1886)	–	*	–	*	–	*	–	–	38
<i>Lecane furcata</i> (Murray, 1913)	–	–	–	*	–	–	–	–	13
<i>Lecane hamata</i> (Stokes, 1896)	–	+	–	+	–	+	–	+	50
<i>Lecane hastata</i> (Murray, 1913)	–	–	–	–	*	–	*	–	25
<i>Lecane hornemanni</i> (Ehrenberg, 1834)	–	–	–	*	–	–	–	–	13
<i>Lecane inermis</i> (Bryce, 1892)	–	+	–	+	–	–	–	+	38
<i>Lecane luna</i> (Müller, 1776)	–	–	–	–	*	*	–	–	25
<i>Lecane pyriformis</i> (Daday, 1905)	–	–	–	–	*	–	–	–	13
<i>Lecane scutata</i> (Harring & Myers, 1926)	*	*	*	*	–	*	*	–	75
<i>Lecane stenroosi</i> (Meissner, 1908)	–	*	–	*	–	–	–	*	38
Lepadellidae									
<i>Colurella adriatica</i> Ehrenberg, 1831	*	*	–	*	*	*	–	–	63
<i>Colurella colurus</i> (Ehrenberg, 1830)	–	*	–	*	–	*	–	–	38
<i>Lepadella patella</i> (Müller, 1773)	–	*	–	+	–	–	–	+	38
<i>Lepadella rhomboides</i> Gosse, 1886)	–	–	–	–	*	+	–	–	25
Synchaetidae									
<i>Polyarthra dolichoptera</i> Idelson, 1925	●	*	o	+	o	+	*	–	88
<i>Polyarthra vulgaris</i> Carlin, 1943	●	–	●	–	o	–	*	–	50
<i>Synchaeta stylata</i> Wierzejski, 1893	o	–	●	–	o	–	*	–	50
Trichocercidae									
<i>Trichocerca pusilla</i> (Jennings, 1903)	+	*	+	*	*	–	*	–	75
<i>Trichocerca ruttneri</i>	–	–	+	–	–	–	–	–	13
Notommatidae									
<i>Cephalodella forficula</i> (Ehrenberg, 1830)	–	o	–	o	–	*	–	–	38
<i>Cephalodella gibba</i> (Ehrenberg, 1830)	–	+	–	o	*	–	–	–	38
<i>Eosphora najas</i> Ehrenberg, 1830	–	–	–	*	–	+	–	–	25
Philodinidae									
<i>Rotaria neptunia</i> (Ehrenberg, 1830)	+	o	o	o	*	o	–	–	75
<i>Dissotrocha hertzogi</i> Hauer, 1939	–	*	–	+	–	o	–	–	38
Mytilinidae									
<i>Mytilina ventralis</i> (Ehrenberg, 1830)	–	+	–	–	*	o	–	–	38

Table 2. (continued)

Species	Sampling Time								F%
	18.05.2022		10.06.2023		03.09.2023		17.12.2023		
	pl	op	pl	op	pl	op	pl	op	
Asplanchnidae									
<i>Asplanchna sieboldi</i> (Leydig, 1854)	○	–	○	–	+	–	+	–	50
Collotheceidae									
<i>Collothecha pelagica</i> (Rousselet, 1893)	–	○	–	○	–	●	–	+	50
<i>Collothecha mutabilis</i> (Hudson, 1885)	–	*	–	+	–	+	–	–	38
Testudinellidae									
<i>Testudinella patina</i> (Hermann, 1783)	*	–	–	–	–	–	–	–	13
<i>Pompholyx sulcata</i> Hudson, 1885	●	–	+	–	○	–	+	–	50
Gastropodidae									
<i>Ascomorpha ovalis</i> (Bergendahl, 1892)	–	*	+	*	+	–	–	–	50
Trochosphaeridae									
<i>Filinia longiseta</i> (Ehrenberg, 1834)	○	*	+	*	+	–	–	–	63
<i>Filinia opoliensis</i> (Zacharias, 1898)	–	–	–	–	*	–	*	–	25
Conochilidae									
<i>Conochilus unicornis</i> Rousselet, 1892	+	–	–	–	–	–	–	–	13
Dicranophoridae									
<i>Dicranophorus epicharis</i> Harring & Myers, 1928	–	+	–	○	–	+	–	+	50
Hexarthridae									
<i>Hexarthra intermedia</i> (Wiszniewski, 1929)	*	–	+	*	–	–	–	–	38
Epiphanidae									
<i>Proalides tentaculatus</i> de Beauchamp, 1907	+	–	–	–	*	–	–	–	25
Cladocera									
Bosminidae									
<i>Bosmina longirostris</i> Müller, 1785	●	*	+	*	–	–	○	–	63
Moinidae									
<i>Moina micrura</i> Kurz, 1875	*	–	+	–	●	–	○	–	50
Chydoridae									
<i>Alona costata</i> Sars, 1862	–	+	–	–	*	○	–	–	38
<i>Disparalona rostrata</i> (Koch, 1841)	–	*	*	–	–	–	–	–	25
Macrothricidae									
<i>Macrothrix laticornis</i> (Jurine, 1820)	–	*	–	*	–	○	–	–	38
Ilyocryptidae									
<i>Ilyocryptus sordidus</i> (Liévin, 1848)	–	*	–	+	–	*	–	–	38
Copepoda									
Cyclopidae									
<i>Cyclops vicinus</i> Uljanin, 1875	○	–	○	–	+	–	–	–	38
<i>Acanthocyclops robustus</i> (Sars G.O., 1863)	○	–	*	–	○	*	+	–	63
<i>Microcyclops rubellus</i> (Lilljeborg 1901)	–	–	–	–	–	+	–	–	13
Ameiridae									
<i>Nitokra hibernica</i> (Brady, 1880)	–	○	–	●	–	–	*	–	38
Laophontidae									
<i>Onychocamptus mohammed</i> (Blanchard & Richard, 1891)	–	*	*	*	*	–	*	–	63
Ergasilidae									
<i>Ergasilus sieboldi</i> von Nordmann, 1832	–	–	*	–	*	–	–	–	25
Lernaeidae									
<i>Lernaea cyprinacea</i> Linnaeus, 1758	–	–	–	–	*	–	*	–	25
Seasonal species numbers	53		54		52		37		

When evaluated according to the frequency index, 3 species ($F \geq 76\%$) were classified as constant, 11 species ($75\% > F \geq 51\%$) were classified as common, 35 species ($50\% > F \geq 26\%$) were classified as occasionally and 17 species ($F < 26\%$) were classified as rare. Among these dense species, *Brachionus quadridentatus* and *Keratella tropica* were determined with the highest frequency (100%) in all seasons. *Polyarthra dolichoptera* (88%), *Brachionus bidentatus*, *Trichocerca pusilla*, *Lecane scutata* and *Rotaria neptunia* (75%), *Brachionus caudatus*, *Lecane bulla*, *Lecane closterocerca*, *Colurella adriatica*, *Filinia longiseta*, *Bosmina longirostris*, *Acanthocyclops robustus*, *Onychocamptus mohammed* (63%), *Brachionus angularis*, *Brachionus calyciflorus*, *Brachionus falcatus*, *Keratella cochlearis*, *Keratella tecta*, *Platylas quadricornis*, *Lecane hamata*, *Polyarthra vulgaris*, *Synchaeta stylata*, *Asplanchna sieboldin*, *Collotheca pelagica*, *Pompholyx sulcata*, *Ascomorpha ovalis*, *Dichranophorus epicharis* and *Moina micrura* (50%) are other zooplanktonic organisms that are frequently seen (Table 2).

DISCUSSION

Temperature increases biological activity in water and accelerates biochemical reactions, which affects the reproduction, feeding and metabolic activities of aquatic organisms. Temperature is one of the most important environmental parameters affecting biodiversity and zooplankton density in aquatic ecosystems (Herzig, 1987; Sharma et al., 2007). Environmental conditions, particularly water temperature, have been reported to have a significant and positive effect on zooplankton diversity and abundance (Rossetti et al., 2009; Dorak, 2013). In this study, where similar results were found, the highest number of species (54 species) was found in the summer season when the average temperature is highest (25.00 ± 0.28).

The amount of dissolved oxygen varies depending on the trophic state of the lakes and the water temperature (Viet et al., 2016). Studies have shown that low oxygen conditions can affect the growth, reproduction and distribution of zooplankton, and it has been reported that dissolved oxygen levels below 5 mg L^{-1} in freshwater can limit the growth of zooplankton (Karpowicz et al., 2020). In the study, the lowest dissolved oxygen was measured in summer and the highest in winter, as high temperature reduces dissolved oxygen in water and increases it at low temperature. The dissolved oxygen values found ($8.05\text{--}9.55 \text{ mg L}^{-1}$) were above 5 mg L^{-1} . Looking at the dissolved oxygen content in the lake, it appears to be suitable for zooplankton life.

Conductivity, an important water quality parameter, is significantly related to zooplankton diversity, abundance and distribution, and an inverse relationship between conductivity and zooplankton species diversity has been

found (Estlander et al., 2009; Tavsanoğlu et al., 2015). In general, conductivity increases in places where there is insufficient water inflow due to evaporation when the water temperature rises. Pollution can increase the conductivity of lakes and rivers, as industrial and anthropogenic effluents often have high conductivity (Wetzel, 1983). While the electrical conductivity in the reservoir was found to be high in summer and fall and low in winter and spring, the conductivity values were found to be at a very low level between 139 and $215 \mu\text{S cm}^{-1}$ and suitable for zooplankton life (Estlander et al., 2009).

pH, which is important for the life cycle of zooplankton, can affect the abundance of zooplankton. Alkaline conditions strongly correlate with primary production and promote zooplankton growth and abundance (Bednarz et al., 2002; Mustapha, 2009), while low pH leads to a decrease in zooplankton abundance, species diversity and extinction of some species (Ivanova and Kazantseva, 2006). In the Surface Water Resources Quality Criteria Regulation (OSİB, 2015), published by the Ministry of Forestry and Water Affairs of the Republic of Türkiye, it is stated that the freshwater pH value, at which many aquatic creatures can sustain their normal life without harming their physiology, should be between 6.00 and 9.00. The pH of the reservoir is between 7.40 and 8.65 and is slightly alkaline and suitable for the survival of zooplankton species (Tessier and Horwitz, 2011).

This study is the first zooplankton study in the young Reyhanlı Reservoir, the construction of which was completed in 2020. 67 species were identified in this study, with the Rotifera being the dominant group with 80.30%.

Various studies have shown that rotifers dominate both qualitatively and quantitatively in stagnant waters, such as most lakes, ponds, reservoirs and wetlands (Jamila et al., 2014; Ismail and Adnan, 2016; Dorak et al., 2019; Saler et al., 2019). In addition, Segers (2007) reported that rotifers occur in almost all types of freshwater habitats, e.g., large permanent lakes, small temporary ponds, intermediate and capillary waters, acidic mineral lakes, fizzy drink lakes, hyperoligotrophic alpine lakes and sewage ponds.

Species reported by various researchers to be good indicators of eutrophic conditions and pollution are *A. coelata*, *B. angularis*, *B. calyciflorus*, *B. quadridentatus*, *B. urceolaris*, *K. cochlearis*, *K. tecta*, *K. tropica*, *L. patella*, *T. pusilla*, *C. forficula*, *R. neptunia*, *C. mutabilis*, *P. sulcata*, *A. ovalis*, *F. longiseta*, *B. longirostris* (Dussart, 1969; Voigt and Koste, 1978; Pesce and Maggi, 1981; Berzins and Bertilsson, 1990; Hansen and Jeppesen, 1992; De Manuel Barrabin, 2000; Petrussek, 2002; Shah and Pandit, 2013; Apaydın Yağcı, 2016; Heneash and Alprol, 2020) was also detected in this study. They also proposed that genus *Brachionus*, *Lecane*, *Trichocerca* and

Keratella can be considered a target taxon for more intensive monitoring of water quality and conservation planning on aquatic environment (Ceirans, 2007). Additionally, *M. micrura*, *C. vicinus* and *A. robustus* are species that are not indicators of eutrophication but are commonly found in eutrophic waters (Jana and Pal, 1985; Crosetti and Margaritora, 1987; Hart, 1990). Unlike these, four species (*P. stylata*, *Collathea pelagica*, *C. mutabilis*, *F. opoliensis*) reported to be oligotrophy indicators by Sládeček (1983) were also found in this study.

The presence of a large number of eutrophication indicator species (17) and a smaller number of oligotrophication indicator species (4) in the dam lake suggests that the dam lake is under the threat of eutrophication, despite being newly established. However, the low abundance of eutrophication indicator species suggests that the dam lake is not currently under the risk of eutrophication. The majority of the species detected are widespread, cosmopolitan species (Keppeler 2003; Keppeler and Hardy 2004; Segers 2007; Melo Júnior et al. 2007; Santos et al., 2013) that have been detected in previous studies in this region (Eastern Mediterranean inland waters).

Some of the species in the study; *A. sieboldii*, *B. angularis*, *B. quadridentatus*, *C. gibba*, *C. pelagica*, *C. adriatica*, *C. colurus*, *F. longiseta*, *K. cochlearis*, *K. quadrata*, *K. tecta*, *L. bulla*, *L. closteroerca*, *L. flexilis*, *L. luna*, *L. pyriformis*, *L. patella*, *P. quadricornis*, *P. dolichoptera*, and *T. pusilla* have been reported to tolerate a wide range of conductivity (Arcifa et al., 1994; De Ridder and Segers, 1997; Baribwegure and Segers, 2001; Pattnaik, 2014). Some others; *B. calyciflorus*, *P. quadricornis*, *L. luna*, *C. adriatica*, *C. colurus*, *C. mutabilis*, *L. patella*, *T. patina*, *Nitokra hibernica* and acidic *K. cochlearis* and *C. gibba* have been reported to be euryhaline (De Smet 1996; De Ridder and Segers 1997; Fontaneto et al. 2006; Jersabek and Bolortsetseg, 2010; Defaye and Dussart 2011). The fewer species in the study are; *B. quadridentatus*, *B. urceolaris*, *L. bulla*, *L. flexilis*, *L. rhomboides* and *R. neptunia* have high tolerance to alkaline waters (De Smet, 1996; Ramdani et al., 2001; Rybak and Bledzki, 2010; De Smet, 1996; Rybak and Bledzki, 2010).

In this study, some species (*L. aculeate*, *L. flexilis*, *L. hamata*, benthic *L. inermis*, *L. stenroosi*, *C. colurus*, *L. patella*, *C. forficula*, *E. najas*, *D. hertzogi*, *C. pelagica*, *C. mutabilis*, *D. epicharis*, benthic *M. laticornis* and *I. sordidus*) was found only on plants. It has been reported by various researchers that the same species are commonly found on littoral plants and to a lesser extent on plankton (Koste, 1978; De Smet, 1993; Hingley, 1993; Segers, 1995; De Manuel Barrabin, 2000; Kuczynska-Kippen, 2000). Therefore, it is thought that the fact that it was found only on plants in the study is due to its general ecological characteristics.

Some species in the study (*A. coelata*, *B. angularis*, *B. budapestinensis*, *B. calyciflorus*, *B. falcatus*, *K. cochlearis*, *K. quadrata*, *K. tecta*, *P. vulgaris*, *S. stylata*, *A. sieboldi*, *P. sulcata*, *F. opoliensis*, *P. tentaculatus*, *M. micrura*, *C. vicinus*, *E. sieboldi*, *L. cyprinacea*) were never found on plants and were only found in plankton. Researchers have reported that these species are widely distributed, mostly found in pelagic, but less commonly on plants and benthic (Hutchinson, 1967; Ruttner-Kolisko, 1974; Margalef et al., 1976; Braioni and Gelmini, 1983; Koste and Shiel, 1986, 1987; Ramdani et al., 2001; Santos et al., 2013).

The crustacean parasitic copepods, *Lernaea cyprinacea* and *Ergasilus sieboldi*, have a life cycle in which males are free-living and adult females go through a free-living stage before becoming parasitic and feed on algae (Molnar and Szekely, 1997; Hossain et al., 2018). Therefore, due to the life cycle of these species, not all stages are parasitic and it is quite normal for them to be found in plankton.

CONCLUSION

The zooplankton species in the dam lake consist of cosmopolitan, widely distributed species that tolerate a wide range of conductivity, salinity and alkalinity. Rotifera was the dominant group, followed by Copepoda and Cladocera. The dominant families were Brachionidae and Lecanidae (Rotifera), Chydoridae (Cladocera) and Cyclopoidae (Copepoda). Although the dam lake is newly established, considering the high number of eutrophication indicator species (17 species).

COMPLIANCE WITH ETHICAL STANDARDS

Conflict of Interest

The author declares that there is no conflict of interest.

Ethical Approval

For this type of study, formal consent is not required. Permission to take water samples was obtained from the Ministry of Agriculture and Forestry with the official letter number E-67852565-140.03.03-5292356.

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Data Availability

The data supporting the findings of this study are available from the corresponding author upon request.

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